

Appendix 1

- 1 **This file includes:**
- 2 Supplementary methods
- 3 Tables A1.1 to A1.6
- 4 Figures A1.1 to A1.9
- 5 Supplementary references

6 **Supplemental Methods:**

7 *Climate projections*

8 A combination of 8 projections were used from 4 different global change models (GCMs) at two
9 relative concentration pathways (RCPs). The RCPs chosen were 4.5 and 8.5, the former
10 representing an emissions-controlled future, while the latter represents an uncontrolled emissions
11 future. The particular combination is based on recommendations from Pierce et al. 2016. The
12 LANDIS model utilizes the following climatological variables: daily precipitation (Figure A1.1
13 and A1.2), daily maximum temperature (Figure A1.3), daily minimum temperature, daily
14 average windspeed, and daily average wind direction that are averaged across the Level II EPA
15 ecoregions in the study area.

16 *Forest succession*

17 NECN (v6.5) simulates both above and belowground processes, tracking C and N through
18 multiple live and dead pools, as well as tree growth (as net primary productivity--a function of
19 age, competition, climate, and available water and N). Soil moisture, as well as movement
20 across the dead pools: wood and litter deposition and decomposition, soil accretion and
21 decomposition are based on the CENTURY soil model (Parton et al. 1983, Scheller et al. 2011).
22 Carbon estimates by pool were validated against Wilson et al. (2013) at the ecoregion level,
23 where the model overestimated total C for only one region but was within one standard deviation
24 for all others (see supplemental Figure A1.4). Forest growth estimates using the climate data for
25 year 2010-2015 for the region were calibrated against the MODIS 17a3 product annual mean for
26 2000 – 2015 (Figure A1.5). Mean landscape value for MODIS was 393 g C/m² (sd 134),
27 while for LANDIS the mean value was 320 g C/m² (sd 312). Reproductive success is
28 dependent on temperature and water.

29 *Fire modeling*

30 The SCRPPLE extension (v2.1) models ignitions by drawing the number of ignitions from a
31 zero-inflated Poisson distribution and allocates them across the landscape with a weighted
32 ignition surface for each type of fire modeled (Scheller et al. 2019). The weather influence on
33 fire is based on the Fire Weather Index (FWI) measures created by the Canadian Fire Prediction
34 System (1992). There are three categories of fires that can be modeled: lightning, accidental
35 (i.e., human started), and prescribed fire. The extension also includes the ability to explicitly set
36 fire suppression effort levels across the landscape as well as by ignition type, where the
37 suppression parameter reduces the probability of fire spread from one cell to another. Effort
38 levels can range from 0 to 3, where 0 is no suppression attempted, to 3 which represents high
39 effort and was designed to mimic current suppression efforts in the Basin (Figure A1.6).
40 However, suppression effectiveness can be limited by weather as well, a maximum wind speed
41 parameter can limit suppression to days only when resources can be deployed safely. That
42 parameter was set at wind speeds of 11 meters per second (~25 miles per hour) in consultation
43 with regional fire personnel. Prescribed fires follow a set of weather prescriptions for when fires
44 can occur (Table A1.2).

45 Contemporary wildfires (2000-2016, from CalFIRE FRAP) were used to parameterize fire
46 spread and size from the Central Sierra Nevada in order to increase the sample size of fires.
47 Mean annual fire area (in ha) for observed data was 117 hectares per year (SD = 309), for

48 modeled data, the mean value was 122 hectares per year (SD = 210). In order to move from fire
49 intensity to fire severity (to encompass the mortality associated with fire), five fire experts
50 working in the LTB provided their estimates of mortality for varying species, age, and intensity
51 combinations. More details about the parameterization of the fire extension are found in Scheller
52 et al. (2019). Suppression effort and fire spread are calibrated at the same time in order to try to
53 account for both forces in recreating the contemporary fire regime.

54 The model calculates three levels of fire intensity, roughly corresponding to flame lengths of: 1)
55 less than 4 ft, 2) between 4 ft. and 8ft., and 3) greater than 8ft. While ignitions are based off of
56 climate, fire intensity is based off of fuel loading within each cell. LANDIS calculates fuel
57 loadings based on the current year's litter, duff, and downed and dead woody debris. When a
58 threshold of fine fuels is exceeded in a cell, the fire intensity increases. This threshold is based
59 off a value of $\sim 1100\text{g/m}^2$ or about 5 tons per acre of fine fuels. The other threshold is based on
60 ladder fuels: a combination of specific species, under a certain age, and over a certain amount of
61 biomass per area, contribute to intensity. Those species contributing to ladder fuels are: Jeffrey
62 Pine, white fir, and incense-cedar, and the cohorts in the cell have to be younger than 40 with a
63 biomass greater than 2000g/m^2 (9 tons per acre). When one threshold is exceeded, fire intensity
64 increases. When both thresholds are exceeded, fire intensity is at its highest. High intensity fire
65 spreads as high intensity fire. In order to try to validate fire intensity for the Basin, the targeted
66 fire intensity value for any of the larger multi-day fires was 40% high, 40% mid, and a 20% low
67 intensity, with high intensity less than 60% of the total fire area. These percentage targets were
68 based on the thematic burn severity values present within the Basin from Monitoring Trends in
69 Burn Severity website.

70 *Insect modeling*

71 A modified version of the Biological Disturbance Agent extension (Biomass BDA v.2.0)
72 (Sturtevant et al. 2009) was used to simulate insect outbreaks for three species of insects: Jeffrey
73 pine beetle (*Dendroctonus jeffrey*), mountain pine beetle (*Dendroctonus ponderosae*), and fir
74 engraver beetle (*Scolytus ventralis*). The extension requires insect-specific resource
75 requirements and assigns a species-specific vulnerability that varies by age. Cells are
76 probabilistically selected for disturbance based upon the species host density at a given site and
77 the presence of non-hosts reduce disturbance probability. The parameters for spread and
78 mortality are outlined in Kretchun et al. (2016), see Table A1.5 and Table A1.6 below. Mortality at
79 an outbreak site is subsequently determined by species' age and host susceptibility probabilities
80 based from empirical field studies (Egan et al. 2010, 2016) and expert opinion, see Table A1.2
81 below. The insects had differing rates of spread per year from previous outbreaks. Mountain
82 Pine Beetle had positive neighbor effects, where pheromones promoted more rapid spread when
83 there were neighboring populations. All insects were able to exploit recently burned stands up to
84 10 years after a fire. Following mortality, dead biomass remains on site and moves to the
85 downed woody debris C pool and the fine woody debris C pool.

86 However, unlike Kretchun et al. (2016), the trigger for an outbreak was changed to be responsive
87 to climate signals. This is because for many beetle species climate influences outbreaks in three
88 ways: low winter temperatures cause beetle mortality; year-round temperatures influence
89 development and mass attack; and drought stress reduces host resistance. Here, we modeled
90 climate influences as a function of drought and mean minimum winter temperature, recognizing
91 that the full suite of climatic influences is necessary for a fully mechanistic model. So long as

92 annual climatic water deficit exceeded a set threshold, in conjunction with mean winter
93 minimum temperatures exceeded a certain threshold, outbreaks could occur. A comparison
94 between the modeled and observed outbreak dataset (USFS Aerial Detection Survey:
95 <https://www.fs.fed.us/foresthealth/applied-sciences/mapping-reporting/index.shtml>) found an
96 overestimation of frequency of occurrence but an underestimation of area impacted by
97 insects (Figure A1.7).

98

99 **Supplemental Tables:**

100 Table A1.1. Suppression effort levels and effectiveness on fire spread

probability.	Fire Weather Index Thresholds		Effort Level		
	Fire Type	Low- mod	Mod- high	Low	Moderate
Accidental	40	60	0	5	10
Lightning	40	60	0	5	10
Rx	40	60	0	0	0

102 Table A1.2. Prescribed fire parameters used for Scenario 5

Prescribed Fire Parameters	
MaximumRxWindSpeed	6.6 (m/s)
MaximumRxFireWeatherIndex	55 (unitless)
MinimumRxFireWeatherIndex	10 (unitless)
MaximumRxFireIntensity	1 (low)
NumberRxAnnualFires	364 (days of year allowable, subject to climate constraints)
FirstDayRxFires	1 (first julian day for allowable fire, subject to climate constraints)
TargetRxSize	72 (hectares)

103

104 Table A1.3. Species parameters used in modeling.

Name	Longevity	Sexual maturity age	Shade tolerance	Fire tolerance	Seed effective dispersal distance (meters)	Maximum dispersal distance (meters)	Vegetative Reproduction Probability	Minimum age veg reproduction	Maximum age veg reproduction	Post-fire regeneration
<i>Pinus jeffreyi</i>	500	25	2	5	50	300	0	0	0	none
<i>Pinus lambertiana</i>	550	20	3	5	30	400	0	0	0	none
<i>Calocedrus decurrens</i>	500	30	3	5	30	1000	0	0	0	none
<i>Abies concolor</i>	450	35	4	3	30	500	0	0	0	none
<i>Abies magnifica</i>	500	40	3	4	30	500	0	0	0	none
<i>Pinus contorta</i>	250	7	1	2	30	300	0	0	0	none
<i>Pinus monticola</i>	550	18	3	4	30	800	0	0	0	none
<i>Tsuga mertensiana</i>	800	20	5	1	30	800	0.0005	100	800	none
<i>Pinus albicaulis</i>	900	30	3	2	30	2500	0.0001	100	900	none
<i>Populus tremuloides</i>	175	15	1	2	30	1000	0.9	1	175	resprout
Non-N fixing, Resprouting	80	5	2	1	30	550	0.85	5	70	resprout
Non-N fixing, Seeding	80	5	2	1	30	1000	0	0	0	none
N fixing, Resprouting	80	5	1	1	30	500	0.75	5	70	resprout
N fixing, Seeding	80	5	1	1	30	800	0	0	0	none

105

106

107 Table A1.4. Harvest removals prescription tables

		Abies concolor	Calocedrus decurrens	Pinus jeffreyi	Abies magnifica	Pinus contorta	Pinus lambertiana	NonnResp	NonnSeed	FixnResp	FixnSeed
Hand Thinning	Age range	1-60	1-64	1-52	1-60	1-73	1-52	10-200	10-200	10-200	10-200
Scenario 1 - 5	Percent removed	-66%	-66%	-66%	-66%	-66%	-66%	-5%	-5%	-5%	-5%
Trees up to 11" dbh	Age range	61-70	65-78	53-68	61-75	74-88	53-64				
	Percent removed	-39%	-39%	-39%	-39%	-39%	-39%				
Mechanical Thinning		Abies concolor	Calocedrus decurrens	Pinus jeffreyi	Abies magnifica	Pinus contorta	Pinus lambertiana	NonnResp	NonnSeed	FixnResp	FixnSeed
Scenario 1, 2, 4, 5	Age range	1-60	1-64	1-52	1-60	1-73	1-52	10-200	10-200	10-200	10-200
Trees up to 24" dbh	Percent removed	-93%	-93%	-93%	-93%	-93%	-93%	-30%	-30%	-30%	-30%
	Age range	61-65	65-71	53-60	61-68	74-80	53-58				
	Percent removed	-70%	-70%	-70%	-70%	-70%	-70%				
	Age range	66-70	72-78	61-68	69-75	81-88	59-64				
	Percent removed	-65%	-65%	-65%	-65%	-65%	-65%				
	Age range	71-75	79-84	69-76	76-82	89-96	65-70				
	Percent removed	-57%	-57%	-57%	-57%	-57%	-57%				
	Age range	76-80	85-91	77-85	83-90	97-105	71-77				
	Percent removed	-45%	-45%	-45%	-45%	-45%	-45%				
	Age range	81-84	92-99	86-95	91-97	106-115	78-83				
	Percent removed	-32%	-32%	-32%	-32%	-32%	-32%				
	Age range	85-89	100-107	96-105	98-104	116-125	84-90				
	Percent removed	-23%	-23%	-23%	-23%	-23%	-23%				
	Age range	90-93	108-115	106-115	105-112	126-136	91-97				
	Percent removed	-17%	-17%	-17%	-17%	-17%	-17%				
	Age range	94-98	116-125	116-126	113-120	137-148	98-104				
	Percent removed	-13%	-13%	-13%	-13%	-13%	-13%				
	Age range	99-103	126-135	127-138	121-127	149-161	105-112				
Percent removed	-8%	-8%	-8%	-8%	-8%	-8%					
Age range	104-108	136-145	139-151	128-135	162-176	113-120					

	Percent removed	-4%	-4%	-4%	-4%	-4%	-4%				
Mechanical Thinning		Abies concolor	Calocedrus decurrens	Pinus jeffreyi	Abies magnifica	Pinus contorta	Pinus lambertiana	NonnResp	NonnSeed	FixnResp	FixnSeed
Scenario 3	Age range	1-60	1-64	1-52	1-60	1-73	1-52	10-200	10-200	10-200	10-200
Trees up to 38" dbh	Percent removed	-95%	-95%	-95%	-95%	-95%	-95%	-30%	-30%	-30%	-30%
	Age range	61-65	65-71	53-60	61-68	74-80	53-58				
	Percent removed	-95%	-95%	-95%	-95%	-95%	-95%				
	Age range	66-70	72-78	61-68	69-75	81-88	59-64				
	Percent removed	-85%	-85%	-85%	-85%	-85%	-85%				
	Age range	71-75	79-84	69-76	76-82	89-96	65-70				
	Percent removed	-85%	-85%	-85%	-85%	-85%	-85%				
	Age range	76-80	85-91	77-85	83-90	97-105	71-77				
	Percent removed	-85%	-85%	-85%	-85%	-85%	-85%				
	Age range	81-84	92-99	86-95	91-97	106-115	78-83				
	Percent removed	-75%	-75%	-75%	-75%	-75%	-75%				
	Age range	85-89	100-107	96-105	98-104	116-125	84-90				
	Percent removed	-70%	-70%	-70%	-70%	-70%	-70%				
	Age range	90-93	108-115	106-115	105-112	126-136	91-97				
	Percent removed	-60%	-60%	-60%	-60%	-60%	-60%				
	Age range	94-98	116-125	116-126	113-120	137-148	98-104				
	Percent removed	-35%	-35%	-35%	-35%	-35%	-35%				
	Age range	99-103	126-135	127-138	121-127	149-161	105-112				
	Percent removed	-20%	-20%	-20%	-20%	-20%	-20%				
	Age range	104-108	136-145	139-151	128-135	162-176	113-120				
Percent removed	-10%	-10%	-10%	-10%	-10%	-10%					
Age range	109-120	146-180	152-240	136-180	177-230	121-160					
Percent removed	-10%	-10%	-10%	-10%	-10%	-10%					
Age range	121-125	181-200	241-252	181-190	231-250	161-180					
Percent removed	-5%	-5%	-5%	-5%	-5%	-5%					

109 Table A1.5. Insect disturbance inputs by insect

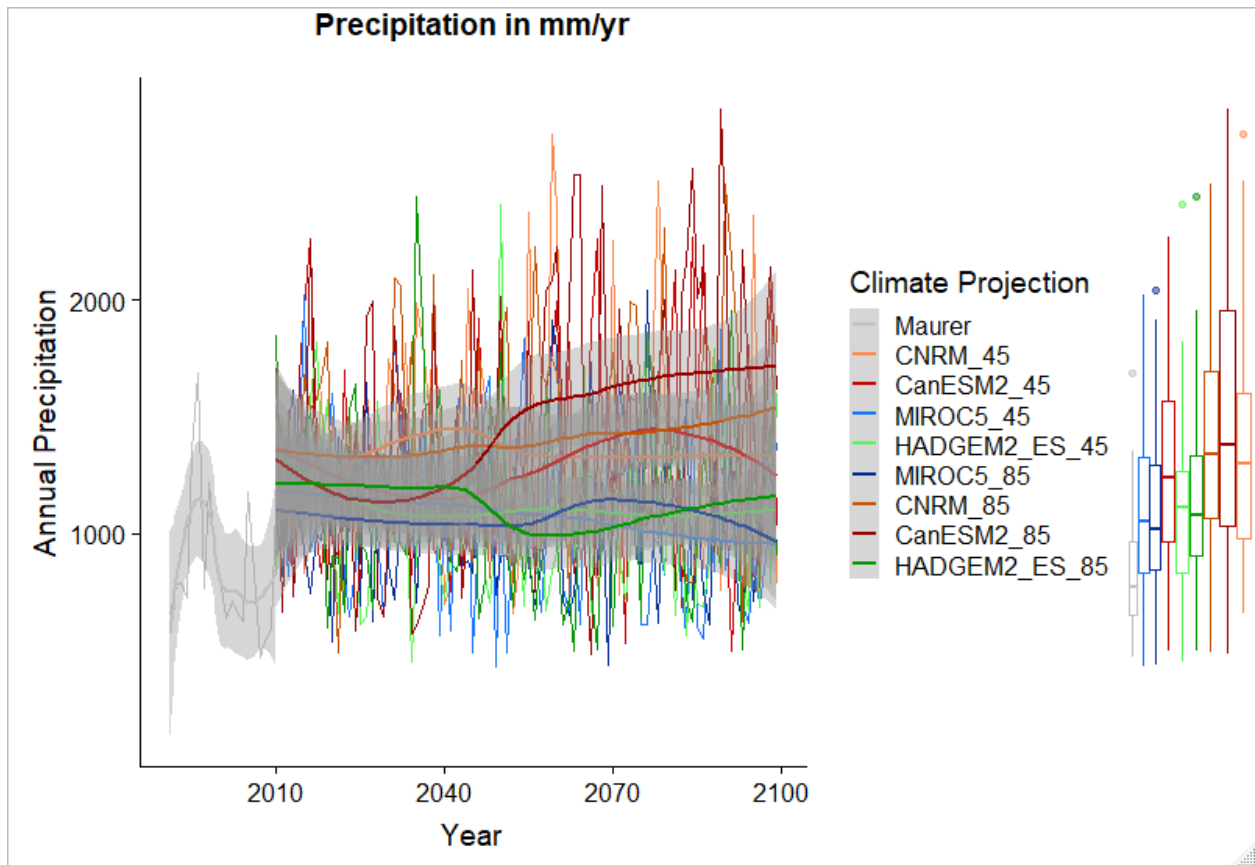
	Fir Engraver		Jeffrey Pine Beetle		Mountain Pine Beetle	
	Parameter	Source	Parameter	Source	Parameter	Source
Dispersal Rate	1000 m/year	Jactel (1991)	600 m/year	Egan (personal comm.)	400 m/ year	Safranik (2006)
Neighborhood Effect	N/A	USFS Fir Engraver Facts (2017)	N/A	N/A	Yes, 2x	Safranik (2006)
Disturbance Modifier	Fire: 100%, 10 years	Schwilk 2006	Fire: 100%, 10 years	Schwilk 2006	Fire: 100%, 10 years	Schwilk 2006

110

111

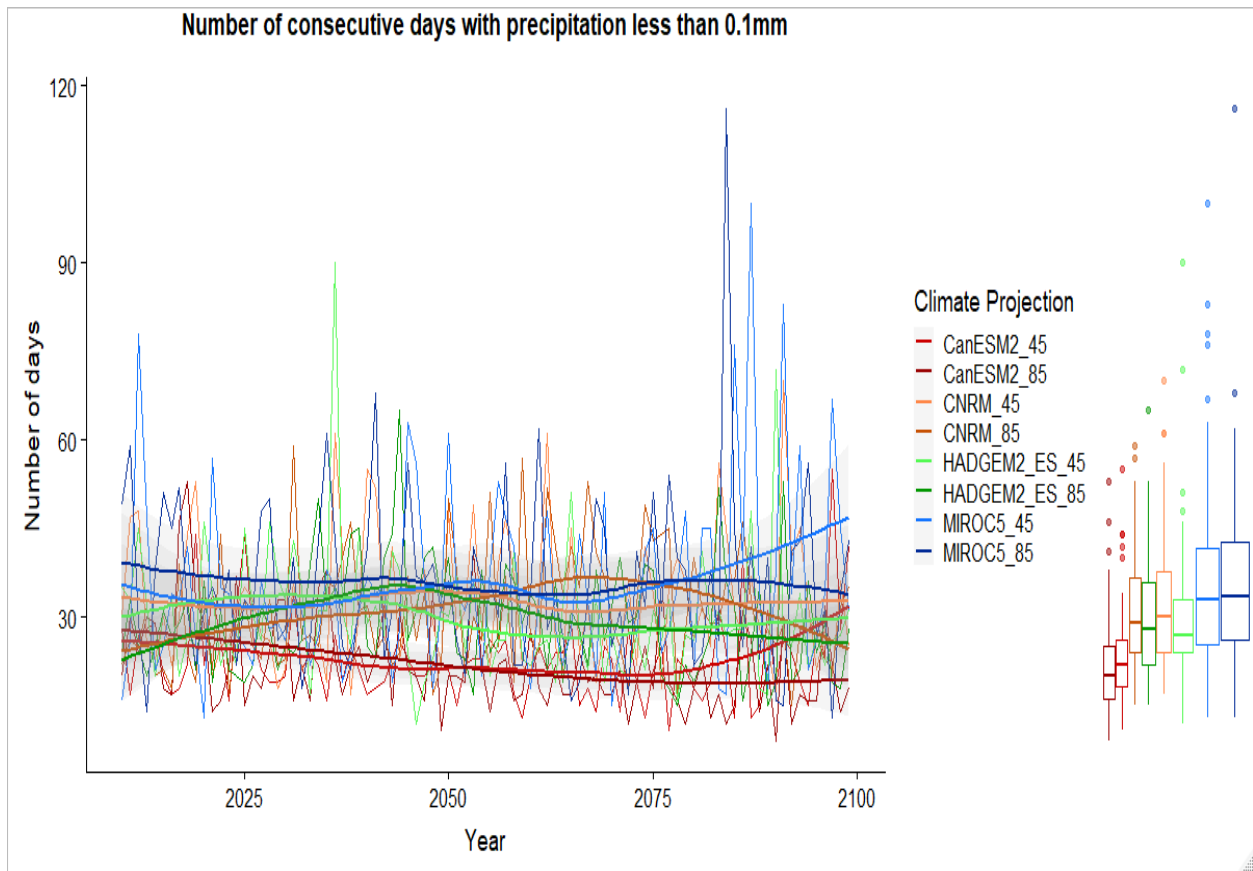
112 Table A1.6: Insect disturbance parameters by insect by host species

	<i>Target Species</i>	Susceptibility			Mortality			<i>Source</i>
		<i>Age Class 1</i>	<i>Age Class 2</i>	<i>Age Class 3</i>	<i>Age Class 1</i>	<i>Age Class 2</i>	<i>Age Class 3</i>	
Fir Engraver	<i>Abies concolor</i>	0-10, 0%	10-60, 65%	60+, 75%	0-10, 0%	10-60, 8%	60+, 12%	Ferrell 1994, Schwilk 2006, Egan (personal comm)
	<i>Abies magnifica</i>	0-10, 0%	10-60, 45%	60+, 55%	0-10, 0%	10-60, 8%	60+, 12%	
Jeffrey Pine Beetle	<i>Pinus jeffreyi</i>	0-20, 10%	20-30, 80%	30+, 80%	0-40, 5%	40-120, 18%	120+, 8%	Egan et al. 2016
Mountain Pine Beetle	<i>Pinus albicaulis</i>	0-20, 33%	20-60, 66%	80+, 80%	0-20, 5%	20-60, 15%	80+, 20%	Safranik (2006), Cole and Amman (1980)
	<i>Pinus lambertiana</i>	0-20, 33%	20-60, 66%	80+, 80%	0-20, 5%	20-60, 25%	80+, 30%	
	<i>Pinus contorta</i>	0-20, 33%	20-60, 66%	80+, 80%	0-20, 5%	20-60, 15%	80+, 20%	
	<i>Pinus monticola</i>	0-20, 33%	20-60, 66%	80+, 80%	0-20, 5%	20-60, 25%	80+, 30%	



115

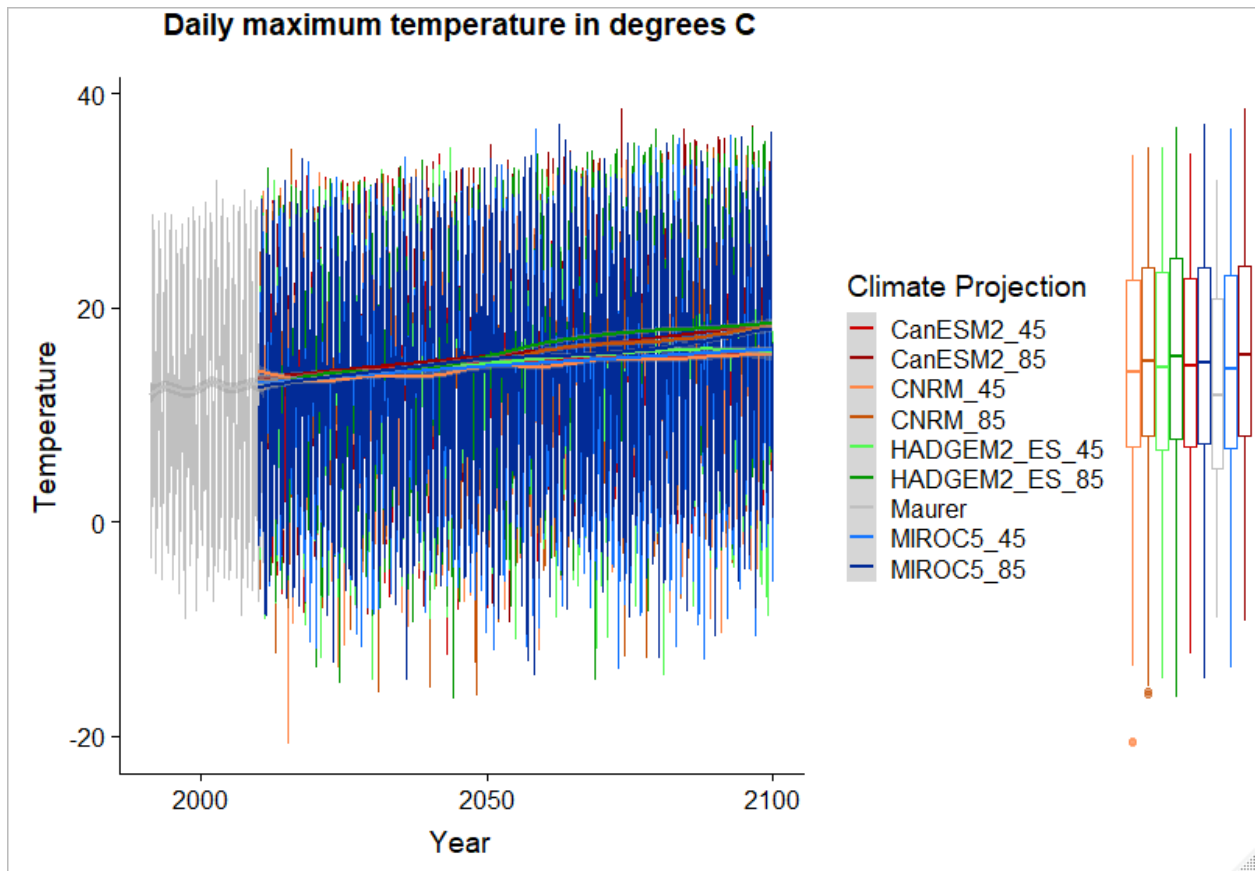
116 Figure A1.1. Projected precipitation in mm yr^{-1} , lines of best fit are GAM estimated, and
117 boxplots represent distribution of annual precipitation for the years 2090-2100.



118

119 Figure A1.2. Projected number of consecutive days with no precipitation, lines of best fit are GAM
 120 estimated, and boxplots represent distribution of consecutive days per year for the years 2090-
 121 2100.

122



123

124 Figure A1.3. Projected daily maximum temperature in degrees C, lines of best fit are GAM
 125 estimated, and boxplots represent distribution of daily temperatures for the years 2090-2100 for
 126 the future climate projections.

127

128

129

130

131

132

133

134

135

136

137

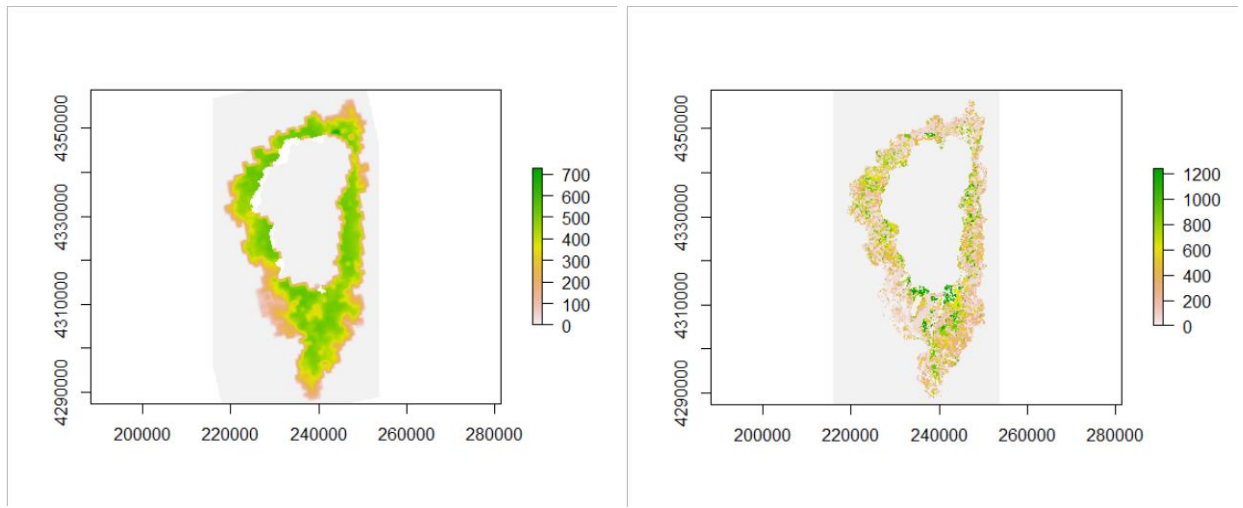


138

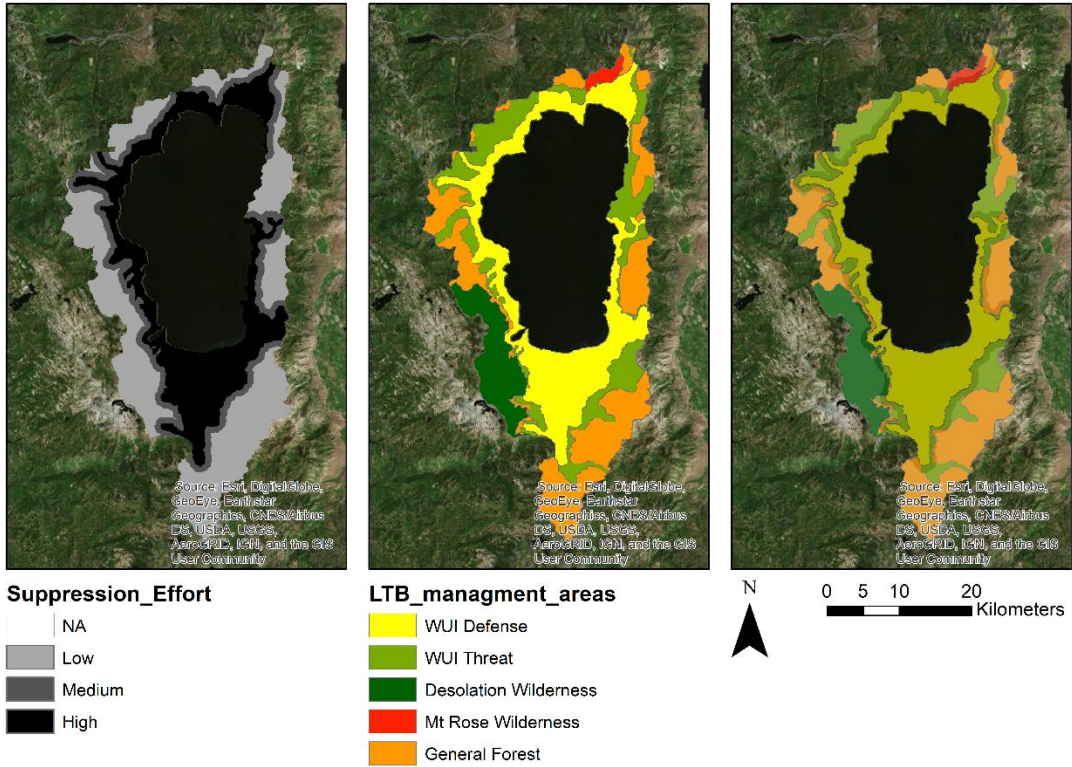
139 Figure A1.4. Observed versus modeled total C, in egagrams C per hectare, by ecoregion, error
 140 m bars represent +/- 1 standard deviation.

141

142 Figure A1.5. Comparison of MODIS (left) and LANDIS (right) estimates of Net Primary
143 Productivity in $g\ C/m^2$. Mean landscape value for MODIS was $393\ g\ C/m^2$ (sd 134), while
144 for LANDIS the mean value was $320\ g\ C/m^2$ (sd 312).

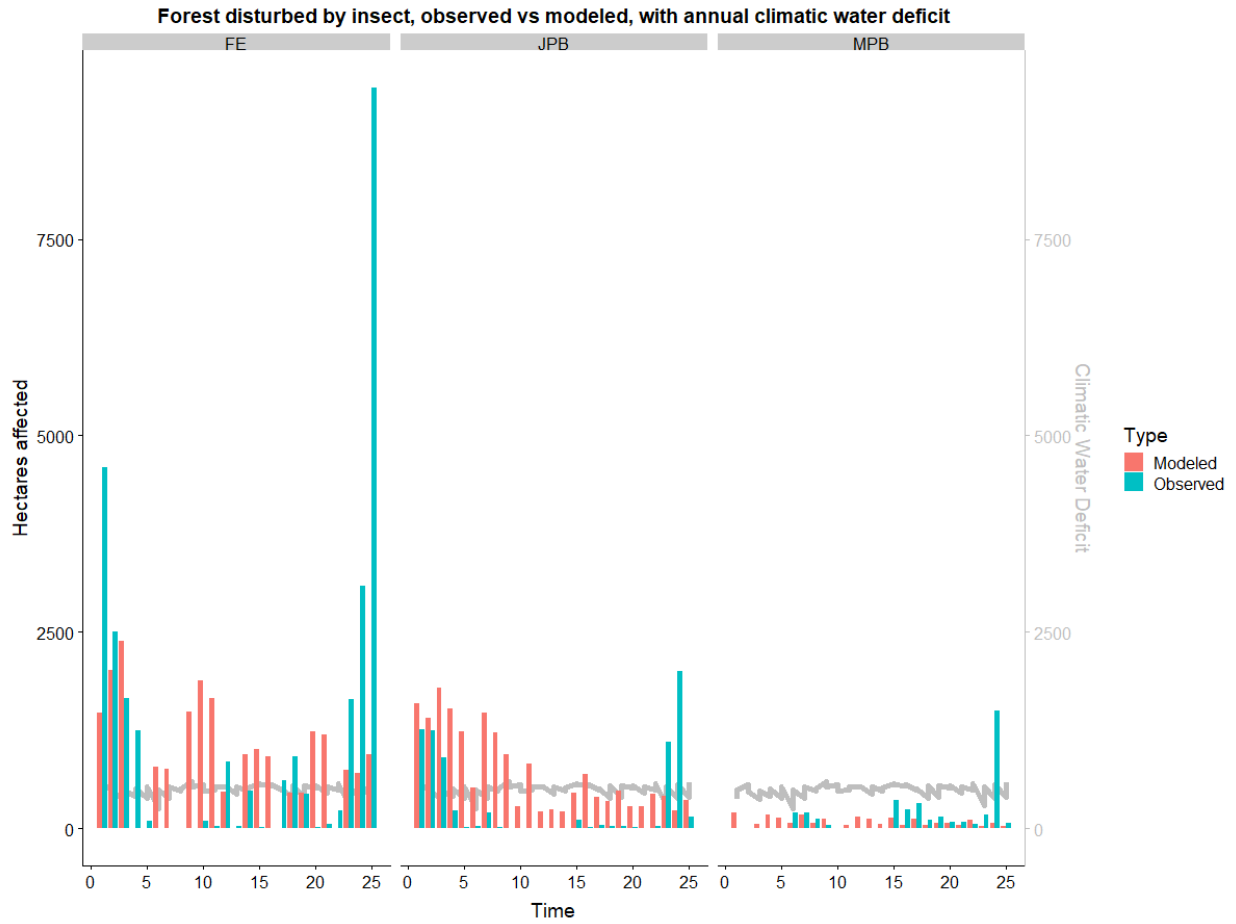


145



146

147 Figure A1.6. Map of suppression (left), management zone (middle), and the overlay of the
 148 effort two (right).

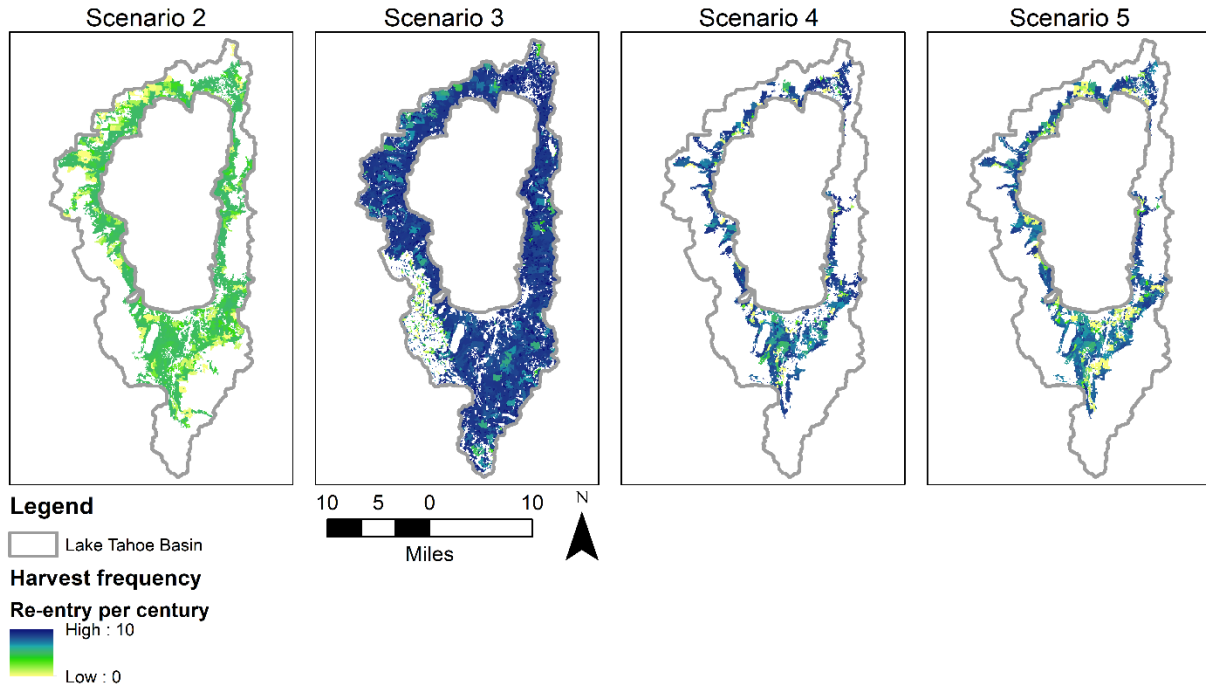


149

150 Figure A1.7. Observed versus modeled number of hectares affected by insect/mortality agent.
 151 Time 0 is equal to 1990, with Time 22-25 corresponding to the 2012-2015 California drought.
 152 FE is fir engraver beetle (*Scolytus ventralis*), JPB is Jeffrey pine beetle (*Dendroctonus jeffreyi*),
 153 and MPB is mountain pine beetle (*Dendroctonus ponderosae*).

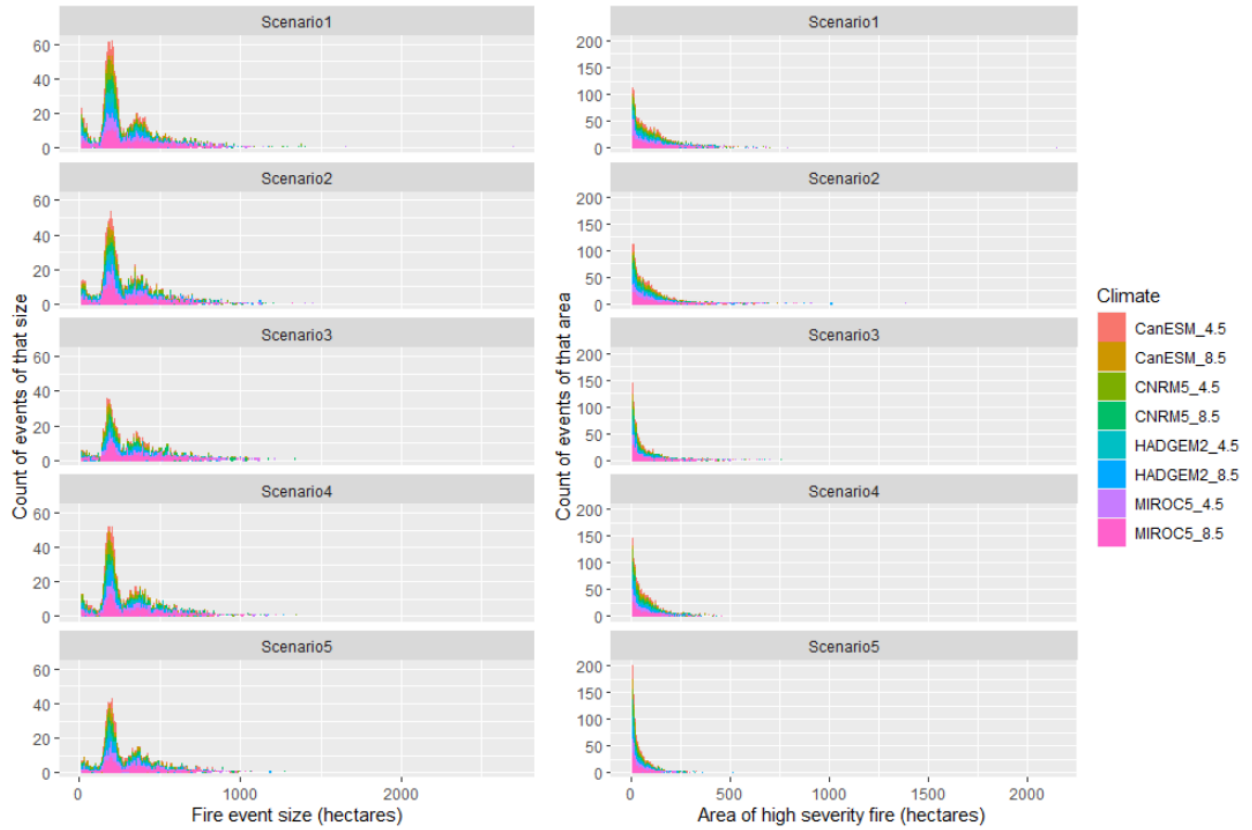
154

155



156

157 Figure A1.8. Harvest return frequency by management scenario. Treatments were
 158 expanded beyond the WUI area in Scenario 3. Scenarios 3 through 5 had a higher
 159 intended treatment frequency.



160

161 Figure A1.9. Histogram of fire sizes (left) and
 162 high severity fire area (right) by scenario and by

high severity fire area (right) by scenario and by

163 **References**

- 164 Egan, J. M., Jacobi, W. R., Negron, J. F., Smith, S. L., & Cluck, D. R. (2010). Forest thinning
165 and subsequent bark beetle-caused mortality in Northeastern California. *Forest Ecology and*
166 *Management*, 260(10), 1832-1842.
- 167 Egan, J. M., Slougher, J. M., Cardoso, T., Trainor, P., Wu, K., Safford, H., & Fournier, D.
168 (2016). Multi-temporal ecological analysis of Jeffrey pine beetle outbreak dynamics within
169 the Lake Tahoe Basin. *Population Ecology*, 58(3), 441-462.
- 170 Ferrell, G. T., Otrosina, W. J., & DeMars, C. J. (1994). Predicting susceptibility of white fir
171 during a drought-associated outbreak of the fir engraver, *Scolytus centralis*, in California.
172 *Can. J. For. Res.* 24: 301-305.
- 173 Hicke, J. A., Meddens, A. J., & Kolden, C. A. (2016). Recent tree mortality in the western
174 United States from bark beetles and forest fires. *Forest Science*, 62(2), 141-153.
175 <http://dx.doi.org/10.5849/forsci.15-086>
- 176 Jactel, H., Van Halder, I., Menassieu, P., Zhang, Q. H., & Schlyter, F. (2001). Non-host volatiles
177 disrupt the response of the stenographer bark beetle, *Ips sexdentatus* (Coleoptera:
178 Scolytidae), to pheromone-baited traps and maritime pine logs. *Integrated Pest Management*
179 *Reviews*, 6(3), 197-207.
- 180 Kretchun, A. M., Loudermilk, E. L., Scheller, R. M., Hurteau, M. D., & Belmecheri, S. (2016).
181 Climate and bark beetle effects on forest productivity—linking dendroecology with forest
182 landscape modeling. *Canadian Journal of Forest Research*, 46(8), 1026-1034.
183 <https://doi.org/10.1139/cjfr-2016-0103>
- 184 Mellen-McLean, Kim, Bruce G. Marcot, Janet L. Ohmann, Karen Waddell, Elizabeth A.
185 Willhite, Steven A. Acker, Susan A. Livingston, Bruce B. Hostetler, Barbara S. Webb, and
186 Barbara A. Garcia. 2017. DecAID, the decayed wood advisor for managing snags, partially
187 dead trees, and down wood for biodiversity in forests of Washington and Oregon. Version
188 3.0. USDA Forest Service, Pacific Northwest Region and Pacific Northwest Research
189 Station; USDI Fish and Wildlife Service, Oregon State Office; Portland, Oregon.
190 https://apps.fs.usda.gov/r6_DecAID
- 191 Parton, W.J., D.W. Anderson, C.V. Cole, J.W.B. Stewart. 1983. Simulation of soil organic
192 matter formation and mineralization in semiarid agroecosystems. In: Nutrient cycling in
193 agricultural ecosystems, R.R. Lowrance, R.L. Todd, L.E. Asmussen and R.A. Leonard (eds.).
194 The Univ. of Georgia, College of Agriculture Experiment Stations, Special Publ. No. 23.
195 Athens, Georgia.
- 196 Pierce, D.W., Cayan, D.R. and Dehann, L., 2016. Creating climate projections to support the 4th
197 California climate assessment. University of California at San Diego, Scripps Institution of
198 Oceanography: La Jolla, CA, USA.
- 199 Safranyik, L.; Carroll, A.L. Pages 3-66 (Chapter 1) in L. Safranyik and W.R. Wilson, editors.
200 The mountain pine beetle: a synthesis of biology, management, and impacts on lodgepole
201 pine. Natural Resources Canada, Canadian Forest Service, Pacific Forestry Centre, Victoria,
202 British Columbia. 304 p. 2006

- 203 Scheller, R., Kretchun, A., Hawbaker, T. J., & Henne, P. D. (2019). A landscape model of
204 variable social-ecological fire regimes. *Ecological Modelling*, 401, 85-93.
205 <https://doi.org/10.1016/j.ecolmodel.2019.03.022>
- 206 Scheller, R.M., Spencer, W.D., Rustigian-Romsos, H., Syphard, A.D., Ward, B.C. and Strittholt,
207 J.R., (2011). Using stochastic simulation to evaluate competing risks of wildfires and fuels
208 management on an isolated forest carnivore. *Landscape Ecology*, 26(10), 1491-1504.
209 <https://doi.org/10.1007/s10980-011-9663-6>
- 210 Schwilk, D. W., Knapp, E. E., Ferrenberg, S. M., Keeley, J. E., & Caprio, A. C. (2006). Tree
211 mortality from fire and bark beetles following early and late season prescribed fires in a
212 Sierra Nevada mixed-conifer forest. *Forest Ecology and Management*, 232(1-3), 36-45.